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Antibiotic-resistant *Escherichia coli* from treated municipal wastewaters and Black-headed Gull nestlings on the recipient river

Martina Masarikova ^{a,b}, Iva Sukkar ^{b,c}, Ivana Jamborova ^b, Matej Medvecky ^b, Ivo Papousek ^d, Ivan Literak ^{b,d}, Alois Cizek ^{a,b}, Monika Dolejska ^{b,c,d,e,*}

^a Department of Infectious Diseases and Microbiology, Faculty of Veterinary Medicine, University of Veterinary Sciences Brno, Brno, Czech Republic

^b Central European Institute of Technology, University of Veterinary Sciences Brno, Brno, Czech Republic

^c Biomedical Center, Faculty of Medicine, Charles University, Pilsen, Czech Republic

^d Department of Biology and Wildlife Diseases, Faculty of Veterinary Hygiene and Ecology, University of Veterinary Sciences Brno, Brno, Czech Republic

^e Division of Clinical Microbiology and Immunology, Department of Laboratory Medicine, The University Hospital Brno, Czech Republic

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ABSTRACT

Wastewaters belong among the most important sources of environmental pollution, including antibiotic-resistant bacteria. The aim of the study was to evaluate treated wastewaters as a possible transmission pathway for bacterial colonisation of gulls occupying the receiving river. A collection of antibiotic-resistant *Escherichia coli* originating both from treated municipal wastewaters discharged to the river Svratka (Czech Republic) and nestlings of Black-headed Gull (*Chroicocephalus ridibundus*) living 35 km downstream of the outlet was obtained using selective cultivation. Isolates were further characterised by various phenotyping and genotyping methods.

From a total of 670 *E. coli* isolates (450 from effluents, 220 from gulls), 86 isolates (41 from effluents, 45 from gulls) showed identical antibiotic resistance phenotype and genotype and were further analysed for clonal relatedness using pulsed-field gel electrophoresis (PFGE). Despite the overall high diversity of the isolates, 21 isolates from both sources showed similar PFGE profiles. Isolates belonging to epidemiologically important sequence types (ST131, 15 isolates; ST23, three isolates) were subjected to whole-genome sequencing. Subsequent phylogenetic analysis did not reveal any close clonal relationship between the isolates from the effluents and gulls' nestlings with the closest strains showing 90 SNPs difference.

Although our study did not provide direct evidence of transmission of antibiotic-resistant *E. coli* to wild gulls via treated wastewaters, we observed gull chicks as carriers of diverse multi-resistant *E. coli*, including high-risk clones, posing risk of further bacterial contamination of the surrounding environment.

1. Introduction

Wastewater treatment plant (WWTP) effluents can serve as one of the most important sources of antibiotic-resistant bacteria for ecosystems including surface waters [1,2]. Rivers receiving the insufficiently treated effluents then represent effective vehicles for spreading of resistant microorganisms over long distances. They are also a medium where bacteria of human and animal origin are brought into contact with naturally occurring bacteria of the aquatic ecosystem, thus enabling the exchange of genetic material and creation of new environmental reservoirs of antibiotic resistance genes [3–5].

Gulls live in close contact with aquatic environments and therefore can become colonised by antibiotic-resistant enterobacteria via treated wastewater effluents [6,7]. *E. coli* resistant to cephalosporins and carbapenems has been previously documented in gulls from a variety of places, e.g. Czech Republic, France, Portugal, Sweden, United States of America, Canada, Australia, Russia or Alaska [8]. Gulls are considered to play a role in the dissemination of intestinal pathogens not only because of their migratory behaviour but also because of their large populations and worldwide distribution [9]. In our previous work, epidemiological relatedness was observed between antibiotic-resistant *Salmonella* spp. isolates originating from wastewater effluents and nestlings of gulls from a breeding colony at recipient river [10]. Here, we broaden the preceding findings with a collection of *E. coli* resistant to beta-lactams and quinolones obtained from the same samples of wastewater effluents and nestlings of gulls. To identify their epidemiological association, and

* Corresponding author at: University of Veterinary Sciences Brno, Palackeho tr. 1946/1, 612 42 Brno, Czech Republic. *E-mail address:* monika.dolejska@gmail.com (M. Dolejska).

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Received 30 August 2023; Received in revised form 20 August 2024; Accepted 20 September 2024 Available online 22 September 2024 2352-7714/© 2024 The Authors. Published by Elsevier B.V. This is an open access article under the CC BY license (http://creativecommons.org/licenses/by/4.0/). hence the possibility of gulls' colonisation with resistant *E. coli* of wastewater origin, the clonal relationship between isolates originating from these two sources was evaluated using antibiotic susceptibility testing, genotyping and whole-genome sequencing (WGS).

2. Material and methods

2.1. Samples collection

Effluents of treated wastewaters were sampled at the discharge from the municipal WWTP (Brno, Czech Republic; 49°8′N 16°38′E) which is flowing into river Svratka. The WWTP processes wastewaters from the city of Brno (approximately 400.000 inhabitants) and surrounding area using a two-stage treatment [11]. A total of 37 effluent samples were collected weekly between March and December 2012 using Moore cotton swabs. Swabs were placed at the effluent for one week and then transported in sterile containers to the laboratory for cultivation.

Chicks of Black-headed Gull (*Chroicocephalus ridibundus*) were onetime sampled in the breeding colony at Nove Mlyny dam ($48^{\circ}53'N$ $16^{\circ}36'E$) situated 35 km downriver of the WWTP in May 2012 simultaneously with bird ringing. It is a period when most of the nestlings in the colony have hatched, but they are not developed enough to fly. A total of 284 cloacal swabs were collected (one sample representing one individual bird) and processed for cultivation.

2.2. Isolation and identification of antibiotic-resistant E. coli

Samples were cultivated in buffered peptone water overnight [10] and subcultured on MacConkey agar (MCA) containing cefotaxime (MCA_{cef}, 2 mg/L) and on MCA with ciprofloxacin (MCA_{cip}, 0.05 mg/L). In the case of gull samples (G), one lactose-positive colony of presumptive *E. coli* isolates was collected. For treated wastewater effluent (WE) samples, where a more diverse spectrum of *E. coli* of different origins can be expected, a maximum of 10 lactose-positive colonies were taken. Obtained colonies were identified by the MALDI-TOF MS method (Microflex LT Biotyper; Bruker Daltonics, Germany). Each WE sample was thus represented by zero to twenty isolates of *E. coli* (0–10 isolates resistant to cefotaxime and 0–10 with reduced susceptibility to ciprofloxacin), while from each G sample zero to two *E. coli* isolates (0–1 isolate colonies) were obtained.

2.3. Antimicrobial susceptibility testing

All *E. coli* isolates were examined by disc diffusion test [12] for susceptibility to 16 antibacterial substances (Table S1). Double-disc synergy test [12] and D68C AmpC & ESBL Detection Set (Mast Diagnostics, UK) were used to evaluate the production of extended-spectrum beta-lactamases (ESBL) and AmpC type beta-lactamases.

2.4. Molecular typing of antibiotic-resistant E. coli isolates

The main criterion for the selection of isolates for further PCR-based typing was a match in the antibiotic resistance phenotypes between at least one WE and G isolate. Based on the profiles of phenotypic resistance, genes encoding selected beta-lactamases, genes for resistance to amphenicols, quinolones, streptomycin, sulphonamides and tetracyclines as well class 1 and 2 integrons were tested by PCR (Table S2). Sanger sequencing was applied to identify specific variants of *bla* and *qnr* genes, and the structure of variable regions of the integrons. PCR was used to assign *E. coli* isolates to phylogenetic groups [13] and for plasmid replicon typing [14].

2.5. Epidemiological relatedness of E. coli isolates

E. coli isolates from different sources with identical resistance profile

and gene set were selected for further epidemiological typing. *XbaI*-restriction in combination with pulsed-field gel electrophoresis (PFGE) was performed according to standardised laboratory protocol [15]. Macrorestriction profiles were analysed with BioNumerics 6.6 fingerprinting software (Applied Maths, Belgium) and isolates were considered to be related if Dice similarity index was \geq 85 % [16]. *E. coli* multilocus sequence typing (MLST) was performed to assign sequence types (STs) (https://mlst.warwick.ac.uk/mlst/). MLST allelic profiles were used to construct a minimum spanning tree employing goeBURST full MLST algorithm as implemented in Phyloviz v2.0 package [17].

2.6. WGS, assembly, and data analysis of related E. coli isolates

The genomic DNA was extracted using NucleoSpin Tissue kit (Macherey-Nagel, Germany), and DNA libraries were prepared using Nextera XT DNA Library Preparation Kit (Illumina, Inc., USA) and subjected to sequencing on the MiSeq instrument (2×250 bp pairedend sequencing). Raw reads were quality trimmed (Trimmomatic v0.39) [18] and *de novo* assembled using SPAdes v3.12.0 [19]. The assemblies were screened for the resistance genes and chromosomal point mutations by ResFinder 3.1 [20], and for presence of plasmid replicons by PlasmidFinder 2.0 [21]. STs and phylogenetic groups of *E. coli* isolates were verified by MLST 2.0 [22] and ClermonTyping tools [23], respectively. ST131 subclones were identified based on the nucleotide sequence of *fimH* using FimTyper 1.0 [24] and variant of *bla*_{CTX-M} [25]. F-type plasmids were assigned to FAB formula using replicon sequence typing [26].

2.7. Comparative genomic analysis of selected isolates

Trimmed reads were mapped against *E. coli* str. K-12 substr. MG1655 (GenBank no. U00096.3) by bowtie2 v2.3.4.1 [27]. Single-nucleotide polymorphisms (SNPs) were detected separately for subsets of isolates belonging to ST131 or ST23, respectively, employing VarScan v2.3.9 [28]. All the sites in which at least one sample had read depth < 8 were discarded from further analysis. SNP distances among isolates were determined using Biostrings library [29] from the R environment v3.5.1. Phylogenetic tree of ST131 isolates based on detected SNP data was built by RAxML v8.2.10 [30] using the maximum-likelihood method. GTR model of nucleotide substitution with Γ rate of heterogeneity estimated from the data was identified as the most appropriate model for analysis through jModelTest v2.1.10 [31]. Robustness of the inferred tree topology was assessed by 500 bootstrap replicates. Final tree topology was visualised via iTOL v5.2 [32].

3. Results

3.1. Isolation and typing of antibiotic-resistant E. coli from wastewater effluents and gulls

A total of 670 *E. coli* isolates were obtained using selective cultivation. In case of WE (n = 450), at least one resistant isolate was obtained from each water sample and multi-resistance (resistance to three or more antibiotic substances) was detected in 66 % and 97 % WE isolates selected on MCA_{cip} (176/265) and MCA_{cef} (180/185), respectively. A total of 181 (64 %, n = 284) individual gulls were colonised by at least one antibiotic-resistant *E. coli* isolate and multi-resistance was demonstrated in 64 % (MCA_{cef}, 81/126) and 57 % (MCA_{cip}, 54/94) of G isolates. ESBL production was detected in 34 % isolates (229/670; 163 WE, 66 G) while 16 % of isolates (108/670; 55 WE, 53 G) produced AmpC type beta-lactamase. For further characterization, 146 isolates (76 WE, 70 G) with the same resistance phenotype were selected for genotyping. Results of phenotypic characterization are summarised in Table S1.

3.2. Detection of selected genetic markers in E. coli to compare WE and G isolates

From the total of 146 *E. coli* with the same resistance phenotype, 86 isolates (41 WE, 45 G) shared the same antibiotic resistance genes (Table S1) and this final set was selected for further genotyping (Fig. S1). Isolates were assigned to various phylogenetic groups with B2 being the most common (31 %, 19 WE, 8 G), followed by B1 (23 %) and A (21 %). Sequencing of the *bla* and *qnr* genes demonstrated the presence of *bla*_{TEM-1} (5 WE, 6 G), *bla*_{CTX-M-1} (2 WE, 3 G), *bla*_{CTX-M-15} (1 WE, 1 G), *bla*_{CTX}. M-27 (9 WE, 2 G), *bla*_{CTX-M-174} (3 WE), *bla*_{CTX-2} (1 WE, 1 G) and *qnrS1* (1 WE, 1 G) variants. A class 1 integron 1.7 kb in size and containing a *dfr17-aadA5* gene cassette was found in six WE and three G isolates. Plasmid replicons were detected in 75 isolates (87 %, *n* = 86) with FIB (71 %) and FIA (24 %) being the most prevalent.

3.3. Phylogenetic relatedness of WE and G E. coli

Thirty-seven different STs were identified among 86 isolates (Fig. S1, Fig. 1) with ST131 being the most common (16 WE, 3 G), followed by ST10 (1 WE, 4 G) and ST23 (2 WE, 3 G). High genetic diversity with 60 unique macrorestriction profiles among 86 isolates was demonstrated using PFGE (Fig. S1). Three subgroups (\geq 85 % similarity) of WE and G

isolates belonging to ST131 (n = 15), ST23 (n = 3) and ST162 (n = 3) were identified (red highlighted in Fig. S1). Isolates of ST131 and ST23 were further typed by WGS (Fig. 2).

All ST131 isolates showed multi-resistance phenotype and carried genes conferring ESBL production. Majority of them (11/15) were classified as H30-R1 lineage carrying $bla_{CTX-M-27}$ and F1:A2:B20 plasmid replicons. ST131 of WE origin (n = 12) were isolated from samples taken at different time points one month apart and all of which preceded the gulls sampling. ST23 included one wastewater isolate sampled two months prior to the two isolates originating from gulls. They were assigned to phylogenetic group C, contained *fimH35* and showed the same resistance phenotype, genotype and plasmid profile. ST162 (2 WE, 1 G) showed 88 % similarity of PFGE profiles, belonged to commensal phylogenetic group B1, were resistant to quinolones and carried FIB replicons. The WE isolates were obtained during two different samplings one month prior to the isolates from gulls.

Precise phylogenetic analysis of *E. coli* ST131 and ST23 was performed (Fig. 2). The number of SNPs detected in *E. coli* ST131 isolates ranged from 77 to 224 but the closest isolates from G and WE showed 90 SNPs difference. *E. coli* ST23 isolates showed the SNPs in range of 222–257. Due to the observed low relatedness among ST131 and ST23 isolates, the wastewater effluents cannot be concluded as a source of contamination of gull's nestlings in this study.



Fig. 1. A minimum spanning tree of 85 E. coli isolates from wastewater effluents and Black-headed Gulls, based on MLST data.

Each ST is represented by a separate circle. The size of the circle depends on the number of isolates. The thickness of the connecting line reflects a number of allelic differences between STs: thick line (1–2 differences), narrow line (3–4), dashed line (5–6). Strain origin: wastewater effluents (WW), gulls (G). One G isolate (R1350e) was not included into the analysis since it could not be assigned to a particular ST.



Fig. 2. Characteristic of 18 E. coli isolates from wastewater effluents and gulls.

^aPhylogenetic tree of *E. coli* ST23 isolates was conducted but not included due to negligible relatedness of the isolates (data not shown). ^bStrain origin: wastewater effluents (WE), gulls (G). ^cAntibiotic resistance genes: cut-off for positive detection was set up for at least 95 % identity and 100 % coverage to reference sequences. ^dChromosomal known point mutations conferring antimicrobial resistance. ^ePlasmid replicons: cut-off for positive detection was set up for at least 95 % identity and coverage to reference sequences.

4. Discussion

The issue of antibiotic resistance in wastewaters has been previously addressed by a number of environmental studies pointing out the insufficient elimination of bacteria during treatment processes [33–35]. Here, the majority of *E. coli* isolated from effluents showed multi-resistance phenotype (79 %, 355/450), and 69 % WE isolates obtained on cefotaxime-supplemented medium (128/185) belonged to ESBL producers. The same municipal effluent was one-time sampled in 2016 followed by cultivation on cefotaxime-supplemented medium and 94 % of obtained *E. coli* (33/35) showed ESBL production [36]. It represents a high increase (from 69 % to 94 %) between the years 2012 and 2016.

In our study, we chose a Black-headed Gull to examine the transmission pathway of antibiotic-resistant enterobacteria into the environment via wastewater effluents. This bird species shows a synanthropic way of life together with a close link to the aquatic environment. Colonised birds can act as reservoirs of antibiotic-resistant bacteria including important human pathogens, therefore they play a significant role in the epidemiology of infectious diseases [37–40].

A large study was conducted in gulls from nine European countries to test the north-south resistance gradient of antibiotic-resistant E. coli [41]. The worst situation was demonstrated in Spain where 61 % of gulls were colonised by E. coli resistant to at least one antibiotic even though antibiotic-supplemented agars were not used to select the isolates. The lowest prevalence of antibiotic-resistant E. coli was observed in gulls nesting in Denmark and Ireland, thus confirming that the north-south resistance gradient as observed in human medicine and livestock production may apply also to wildlife. In our study, the level of antibioticresistant E. coli was even higher (64 %) than in Spain, however, this result should be taken with caution as we used a different isolation method with antibiotic-supplemented agars. However, an isolation method with antibiotic-supplemented agars was used in another Swedish study [42] to evaluate gulls' colonisation by ESBL-producing E. coli in Malmö and Gothenburg with a final incidence of 17 %. High occurrence of antibiotic-resistant E. coli including ESBL producers in gulls in the Czech Republic has been reported previously using antibiotic-free media [43,44]. Up to 36 % of gulls from four different colonies located in northern and southern Moravia, Czech Republic, carried antibiotic-resistant E. coli. One of these four colonies was the same (Nove Mlyny colony) as in the present study and the occurrence of antibioticresistant E. coli in gulls was there 26 % in 2005 [43]. As antibioticsupplemented media was used in the present study to select for resistant isolates, the lower carriage rate in 2005 can be explained by the different selection approach rather than the increasing prevalence of antibiotic resistance in gulls.

E. coli ST131 clone dominated and was identified in both sources. This clonal group is clinically important, globally disseminated and dominates in the aetiology of human extraintestinal coli-infections [45]. In addition to its ability to induce serious diseases, it commonly shows production of ESBL encoded by *bla*_{CTX-M-15} and *bla*_{CTX-M-27} and resistance to fluoroquinolones [46,47]. Interestingly, $bla_{\text{CTX-M-27}}$ was found in eleven ST131-H30R1 isolates from both sources (WE, G) while only one isolate ST131-H30Rx with $bla_{CTX-M-15}$ was identified. Of note, $bla_{CTX-M-15}$ 27-positive isolates belonging to the H30R1 clone have been isolated primarily from various clinical materials in a hospital located in the city of Brno [48]. Wastewaters from this hospital are treated by the same WWTP that was examined in our study. Other clinically important clonal lineages, namely ST69 and ST95, were detected in wastewater effluents and gulls in our study. These STs are frequently associated with neonatal meningitis, bloodstream and urinary tract infections in humans and have also been reported in various non-human sources such as wildlife, pets, or meat from retails [49].

The clonal relationship between enterobacteria isolated from wastewaters and water birds has recently been studied [7]. Extensive genetic relatedness through the PFGE and MLST methods was demonstrated on a set of quinolone-resistant *E. coli* isolated from clinical cases of human infections and from various environmental sources. In particular, isolates with similar genotypes originating from hospitalised patients, raw hospital wastewaters, urban surface water, treated wastewaters and gulls were identified. In another study from Bangladesh, *E. coli* isolates resistant to beta-lactams and belonging to the same STs commonly colonising humans and poultry were found in gull faeces [50]. We also observed a similar antibiotic profile, same STs and relatedness by PFGE of isolates originating from effluents and gulls, however the use of high-accurate WGS did not confirm the genetic identity across these two sources but rather showed some level of phylogenetic relatedness for some isolates.

5. Conclusion

Our work focused on the comparison of antibiotic-resistant *E. coli* isolates from wastewater effluents and gulls. We have shown that treated wastewaters discharged into the river Svratka contain high-risk multi-resistant *E. coli* clones despite the previous treatment process in the municipal WWTP. Our study also demonstrated a high carriage rate of antibiotic-resistant *E. coli* in Black-headed Gull nestlings in the colony downstream the river Svratka. In both sources, a high-risk ST131 clone was frequently detected, pointing to the probable circulation of clinical

isolates in the environment.

In our collection of *E. coli*, we did not observe any closely related strains originating from these two sources thus we cannot conclude if contaminated river water can be a vehicle contributing to the colonisation of wild birds. We are aware of the limitation in the study design as a large preselection of molecularly typed isolates and the subsequent low number of sequenced isolates. Further work focused on other possible bacterial sources including feeding habits of gulls is required to clarify sources of contamination of the gulls by antibiotic-resistant bacteria.

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CRediT authorship contribution statement

Martina Masarikova: Writing – review & editing, Writing – original draft, Methodology, Investigation, Funding acquisition, Conceptualization. Iva Sukkar: Writing – review & editing, Writing – original draft, Investigation, Data curation. Ivana Jamborova: Writing – review & editing, Writing – original draft, Methodology, Investigation. Matej Medvecky: Writing – review & editing, Writing – original draft, Visualization, Data curation. Ivo Papousek: Writing – review & editing, Writing – original draft, Investigation. Ivan Literak: Writing – review & editing, Writing – original draft, Supervision, Project administration, Conceptualization. Alois Cizek: Writing – review & editing, Writing – original draft, Supervision, Project administration. Monika Dolejska: Writing – review & editing, Writing – original draft, Validation, Supervision, Resources, Project administration, Funding acquisition, Formal analysis, Conceptualization.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Data availability

The assemblies are available under BioProject PRJNA1007571, accession numbers SAMN37074019- SAMN37074036.

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